

## Review

# Bisphenol A and its Analogues: Human Exposure and Biological Effects—A Review

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**Abstract.** The most prevalent exposure of bisphenol A (BPA) to biological systems has directed health organizations to reduce its safe dosage and the stringent regulation by developed countries that led to ban of BPA-based (Bisphenol analogues) baby bottles. This in turn has opened a new window for ongoing entry of bisphenol analogues into consumer market. Recently, there has been tremendous growth in both production and application of bisphenol analogues to achieve the “BPA-free” claim. Nevertheless, due to growing epidemiological evidence concerning the toxicological effects of these bisphenol analogues in both *in vivo* and *in vitro* systems, the debate regarding the safety concerns over bisphenols is back. Structural analogues of bisphenol A had been identified in food products, human and environment matrices. Present review is an attempt to recapitulate the presence of bisphenols in food and environment matrices as well as their concerning physiological effects in animal models and human groups. But, due to structural analogy of these substitutes, their endpoints on biological functions are comparable to original compound or in certain situations, more harmful than original compound. Unfortunately, other potentially harmful alternatives are emerging and it is therefore advised that the replacement of bisphenol A with other structural analogues must be executed with great care.

**Keywords:** bisphenol analogues, toxicity, food, environment, animal studies, human groups

## Introduction

Bisphenol A, a man-made plastic monomer, virtually present everywhere in industrial world and human exposure to this compound is mainly through dietary and non-dietary sources. Epidemiological studies have evinced that major cause of neuro-behavioural diseases might be due to human exposure to endocrine disrupting chemicals such as bisphenol A (Ikhlas and Ahmad, 2020).

In recent years, scientific knowledge and public awareness has been increased concerning the negative endpoints of BPA (Vandenberg *et al.*, 2013; Chapin *et al.*, 2008). Owing to endocrine disrupting and putative obesogenic endpoints of BPA, the use of its analogues is growing (Zhu *et al.*, 2020). Furthermore, consumer requirement for “bisphenol A-free” products resulted in substitution of BPA with other related chemicals as stated previously by Vandenberg *et al.* (2017) and Mathew *et al.* (2014). In this regard, BPS (4-hydroxyphenyl sulfone), BPF (4,4'-methylene diphenol)

and BPAF (4,4' hexafluoro isopropylidene) diphenol) are major alternatives of BPA in manufacturing polycarbonate (PC) plastics as well as epoxy resins (Dietrich *et al.*, 2017; Liu *et al.*, 2017) and bisphenol AF (BPAF).

On the other side, the safety of these potential alternatives regarding the induction of endocrine disrupting activity in human and animals is not ascertained absolutely (Rosenfeld, 2017). Their potential endpoints by genetic and non-genetic ways may have fostered the substitution of BPA with bisphenol analogues, which unfortunately have the same ill-effects. Several analogues and their derivatives have already been reported in human and environment and there is a growing concern over low dose and matrix related effects of these compounds (Caballero-Casero *et al.*, 2016).

Among these, BPB, BPF, BPS and other analogues have been recognized as emerging contaminants in food, beverages (Cunha and Fernandes, 2013; Liao and Kannan, 2013; Gallart-Ayala *et al.*, 2011), environmental media (Yang *et al.*, 2014 a, b) and biological samples (Zhou *et al.*, 2014). Many studies have reported the

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occurrence of BPS and BPF in several personal care products and paper products (Liao and Kannan, 2014 a,b; Liao and Kannan, 2013; Liao *et al.*, 2012 a). Strikingly, occurrence of BPF in mustard also recognizes it as natural compound of food (Zoller *et al.*, 2016). This overview is to shed light on studies reporting toxicological impact of BPA and its analogues in *in vivo* (animal studies) and human groups. Also, to draw the attention of regulatory agencies to confront the burgeoning use of BPA and its analogues in consumer products.

**Applications of bisphenol analogues.** Epoxy coatings are commonly used in food containers and in drinking water supplies, water pipes in home and other buildings. Main constituents of epoxy coatings are bisphenols like BPF, BPA and their reactive polymers, bisphenol F diglycidyl ether (BFDGE) and bisphenol A diglycidyl ether (BADGE), respectively. Therefore, it could be possible that leachates from epoxies might be susceptible to chemical disinfection *i.e.*, free chlorine and monochloramine. Earlier studies have suggested that BADGE may migrate into drinking water and undergo hydrolysis (Lane *et al.*, 2015 a and b). BPF is widely used as liners, lacquers, water course and food container coatings (Cabaton *et al.*, 2009). Furthermore, the applications of BPAF as a cross-linker in optical fibers, fluoroelastomers and a standard monomer for polyesters (Konno *et al.*, 2004), polyamides and polycarbonate copolymers have been reported earlier (Baradie and Shoichet, 2005).

Industrial applications of BPS has been recognized for electro plating solvent, phenolic resins (Clark, 2012) and products sold as “BPA-free paper”. BPS in consumer products like can coatings, epoxy glues and food-contact paper products, is highly common these days (Mathew *et al.*, 2014; Naderi *et al.*, 2014). A number of dealers switched to its analogue BPS, hypothetically considering the later to be safer (Russo *et al.*, 2017) hence, increasing production of BPS per annum (Rochester and Bolden, 2015). Since, BPS has considerably lower estrogenic effects, good stability against elevated temperature, it was deliberated a “safe BPA alternative” as reported previously (Kuruto-Niwa *et al.*, 2005). Nevertheless, due to negative endpoints pertaining to exposure to bisphenol A as well as other BPs like BPS and BPF (Zhou *et al.*, 2014), are also under scrutiny by different world health organizations.

**Biological effects of bisphenol analogues.** From viewpoint of reproductive toxicology, the replacement of BPA, an extensively used constituent of plastic and non-plastic materials, with its analogues like BPS seems most important challenge (Žalmanova *et al.*, 2016). Several publications based on *in vitro* studies delved into strong estrogenic (Rosenmai *et al.*, 2014; Teng *et al.*, 2013) and anti-androgenic effects (Eladak *et al.*, 2015; Molina-Molina *et al.*, 2013). On the basis of *in vitro*, many researchers (Feng *et al.*, 2016; Peyre *et al.*, 2014; Salvesen and Walsh, 2014; Fic *et al.*, 2013; Lee *et al.*, 2013) have showed that BPA may be cytotoxic and elicit DNA damage. It has been established that either BPA induced or was associated with ovary disorders in laboratory animals and human epidemiological studies (Peretz *et al.*, 2014).

Earlier studies indicated estrogen effect of BPS in rats (Owens and Ashby, 2002) through a postnatal exposure both at low and high dosage levels (Yamasaki *et al.*, 2004). Likewise, investigation by Vinas and Watson (2013a) and (2013b) established equal or greater estrogen efficacy compared to estradiol, BPS possessed the ability to stimulate the membrane receptor pathways usually up-regulated by estradiol. However, an *in vitro* study conducted by Rochester and Bolden (2015) pointed out a weak estrogen effect of BPS than activity established by estradiol. Therefore, regulation of BPF is uncertain due to its chronic intake by majority of people (Dietrich and Hengstler, 2016). Recently, Andújar *et al.* (2019) and Pelch *et al.* (2019) have reviewed the health consequences of BPA analogues particularly related to obesity and other physiological effects.

**Animal studies.** The effect of BPA and its derived compounds (3-hydroxybisphenol A and bisphenol A 3, 4-quinone) was examined on rat performance. It was noticed that only one intracisternal administration of BPA into Wistar male rats (5-day-old) elicited considerable hyperactivity in rats (4-5 weight), which was almost 1.3 times active during nocturnal stage compared to control. Conversely, its metabolites at similar quantity (87 nmol) did not increase the motor activity. The analysis of treated brain through GC-MS have shown that 7% of parent compound resided in brain at 2 months however, its metabolites were not observed (Ishido *et al.*, 2011). In research work of Abdelhaffez *et al.* (2017), the exposure of BPA was investigated on lung tissues of male rats. It was reported

that chronic exposure might contribute in pulmonary inflammatory maladies with likely the initiation of lung fibrosis. The compendium of various studies reporting physiological effects *in vivo* of bisphenol analogues have been described in Table 1.

It has been reported earlier that BPS exposure was found to elicit acute toxicity in *Daphnia magna* while, postnatal exposure led to uterine development in rats representing negative health effects of BPS *in vivo* (Yamasaki *et al.*, 2004; Chen *et al.*, 2002). Singh *et al.* (2016) studied the effect of BPA on male reproduction in Kadaknath chicken through oral BPA doses *i.e.*, 1 or 5 mg/Kg body weight for a period of 7 weeks. The birds with BPA exposure (1 mg/Kg body weight) had highest semen volume than other groups while, the birds with BPA exposure (5 mg/Kg body weight) most considerably reduced sperm concentration compared to other treatment and control. The study outcomes show that bisphenol A indicated adverse effect on characteristics of sperm in chicken without impacting fertilization potential.

Past authors (Qiu *et al.*, 2016, 2015; Kinch *et al.*, 2015; Ji *et al.*, 2013) substantiated the hypothesis that this compound may disrupt both reproduction and developmental effects in zebrafish thru endocrine ways. It has also been evinced that exposure to BPS may affect body weight as well as neuro behaviors in male offspring exposed developmentally (Kim *et al.*, 2015). Published data have established the developmental toxicity of BPA in a range of models. Recently, the developmental toxicity of BPA has been evaluated during three developmental stages of zebrafish and findings showed that BPA exposure reduced reproductive development though, most significant changes were reported in treatment group having lowest concentration. Genetic effects on gamete development might be the cause of secondary effects of decreased fertilization, embryonic mortality and abnormalities (Chen *et al.*, 2017).

In another study, fish (*Labeo rohita*) were administered to a sublethal BPA concentration of 0.73 mg/L for 35 days and hemato-biochemical and enzyme assays were done at the end of each week. The median lethal BPA concentration for *L. rohita* was assessed to be 7.3 mg/L for 24 h. The data evinced a significant ( $P < 0.05$ ) reduction in hematological and biochemical factors in

BPA treated *L. rohita* than control groups (Krishnapriya *et al.*, 2017). In a research work of Little and Seebacher (2015), they speculated that BPA toxicity could differ with environment temperature. For the purpose, zebrafish (*Danio rerio*) was exposed to an environmentally related and artificially higher BPA concentrations at two acclimation temperature. They reported that the environmentally relevant level of BPA (20 µg/L) impaired swimming activity, heart rate and gene expression.

Faheem *et al.* (2017) probed the impact of BPA exposure (10, 100 and 1000 µg/L) on thyroid and steroid hormone levels of a cyprinid (*Catla catla*). They found that plasma estradiol considerably augmented as a result of BPA exposure (@ 100 and 1000 µg/L). After two weeks, a substantial reduction in plasma testosterone, triiodothyroxine and thyroxine was documented. Moreover, plasma FSH levels displayed a significant escalation only at BPA exposure of 10 µg/L. BPA exposure at 1000 µg/L significantly increased the plasma luteinizing hormone in fish. Hence, the alteration in sex hormone as well as gonadotropin levels can elicit reproductive dysfunction.

Nevertheless, BPA and its halogenated derivatives (H-BPAs) have been identified in organisms and environment but, there is limited data regarding their toxicity, particularly chronic exposure at low doses. In this context, Huang *et al.* (2017) systematically evaluated and compared the impact of BPA and H-BPAs on reproduction and growth effects of *Oryzias melastigma* at various development stages.

Bisphenol A and its derived compounds such as tetrachlorobisphenol A (TCBPA) and tetrabromobisphenol A (TBBPA) were observed to escalate embryonic heartbeat. Also, BPA exposure slowed down oocyte development in ovary, decreased estrogen and testosterone levels rather than HBPAs. While, in male fish, insignificant change was noticed in testis. However, BPA had no substantial effect on hatch time whereas, TBBPA and TCBPA reduced hatching rate. Overall, TCBPA exhibited most toxic effect on hatching followed by TBBPA and BPA.

**Human studies.** Biomonitoring studies divulge that human BPS exposure is prevailing (Ye *et al.*, 2015). Similar to BPA, human are exposed to BPS primarily by dermal absorption and ingestion from dust, can

**Table 1.** Toxicological effects of BPA and bisphenol analogues in *in vivo* (animal studies)

Compounds	Model organism	Exposure doses	Exposure duration	Biological effects	References
BPAF	Adult male rats	2, 10, 50 and 200 mg/Kg/day	14 days	A reduction in testosterone was probably a direct effect of bisphenol F exposure to testis.	(Feng <i>et al.</i> , 2012)
BPS	Zebrafish	0.1, 1.0, 10 and 100 µg/L	75 days	Together, developmental BPS exposure at low concentrations adversely affected several parts of endocrine system in Danio rerio.	(Naderi <i>et al.</i> , 2014)
BPA, BPS and BPF	Juvenile female rat	10 µg/Kg bw/day	From GD 12 to parturition	BPA, BPS and BPF differentially affected 5α-reductase and genes associated with dopamine and serotonin in female prefrontal cortex.	(Castro <i>et al.</i> , 2015)
BPA	Rat	0.5 and 50 mg/Kg bw/day	Through out lactation until weaning	Neonatal BPA exposure significantly enhanced oxidative stress, reduced antioxidant enzyme activity and caused damage to DNA. BPA elicits long-term effects on liver that led to toxic effects on liver of female rat offspring.	(Eid <i>et al.</i> , 2015)
BPA	Zebrafish embryo	0.0068 µM	9, 12, 24, 36 and 48 hpf	A very low BPA dose caused 180% increases in neuronal birth/ neurogenesis in hypothalamus.	(Kinch <i>et al.</i> , 2015)
BPS	Zebrafish embryo	0.0068 µM	9, 12, 24, 36 and 48 hpf	Strikingly, exposure of embryonic zebrafish with BPS caused 240% increases in neurogenesis within hypothalamus. Taken together, data suggest that both BPA and BPS impact the hypothalamic growth and may possibly act through the same AroB-mediated mechanism.	(Kinch <i>et al.</i> , 2015)
BPAF	Zebrafish	5, 25 and 125 µg/L	From 4 hpf to 120 dpf (from embryo to adult)	Parental exposure to BPAF at environment relevant concentration tend to delay hatching of offspring but, possible adverse effects of BPAF in offspring need further investigation.	(Shi <i>et al.</i> , 2015)
BPAF	Zebrafish larvae	5, 50 and 500 µg/L	72 to 168 hpf	BPAF exposure caused toxicity of thyroid endocrine. Though, effective BPAF concentrations were much greater than common environment.	(Tang <i>et al.</i> , 2015)
BPA	Rat	50 µg/Kg bw/day	From 15 to 30 days post-partum	A differential effect of BPA and exposure was demonstrated on spermatogenesis establishment in pre-pubertal rats and also a toxic effect on blood-testis-barrier establishment.	(Brouard <i>et al.</i> , 2016)
BPAF	Rat	100 mg/Kg bw/day	From GD 3-19	Gestational as well as lactational BPAF exposure in mother could harm reproductive functioning in male offspring.	(Li <i>et al.</i> , 2016)
BPS	Mice	0.2, 1.5 and 50 µg/Kg bw/day	From GD 0 to offspring at 23-weeks old	BPS was designated as obesogenic at lower dosages and after perinatal and chronically exposed male mice.	(Moral <i>et al.</i> , 2016)
BPS	Rat (male)	1-50 µg/Kg day	For 28 days	Treatment of male rat with BPS prompted oxidative stress likewise, it also heightened antioxidant enzyme activity in tissue. BPS tend to induce oxidative stress in testis; may impact rat spermatogenesis.	(Ullah <i>et al.</i> , 2016)
BPA	Rat	0.5 and 50 µg/Kg day	During gestation and breast-feeding	Animals exposed to BPA established that folliculogenesis and steroidogenesis are the targets of bisphenol A within ovary.	(Santamaria <i>et al.</i> , 2016)
BPS	CD-1 Mice	2 or 200 mg/Kg/d	During pregnancy and lactation	Mother are susceptible during pregnancy and lactation; also the developmental exposure to BPS alters maternal behavior later in adulthood.	(Catanesi and Vandenberg, 2017)
BPAF	Mice	0.4, 4.0 mg/Kg	During pregnancy	Maternal exposure to BPAF considerably impact emotion-related behavior in adolescent mice offspring. Moreover, male offspring with a great possibility to have symptoms of anxiety, depression and may suffer memory impairment.	(Gong <i>et al.</i> , 2017)
BPS	CD-1 Mice	200 µg/Kg/day	From GD 8 to postnatal day 19	In female mice, BPS changed the expression of estrogen-responsive genes in uterus as well as ovary. However, BPS-exposed female mice responded abnormally to an additional dose of estrogen, showing enhanced response in uterus but, decreased responses in ovary.	(Hill <i>et al.</i> , 2017)

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Compounds	Model organism	Exposure doses	Exposure duration	Biological effects	References
BPA and BPF	Female C57BL/6 mice	10 mg/Kg bw	From GD 11.5 to 18.5	Treatment of female mice with BPF significantly altered offspring behavior causing an increase in anxiety and depression. Also, the impact of BPF was stronger compared to BPA.	(Ohtani <i>et al.</i> , 2017)
BPS	Male rats	30, 60 and 120 mg/Kg BW/day	30 days	BPS may impair blood functions and increase cardiovascular menaces in rats.	(Pal <i>et al.</i> , 2017)
BPA	C/D-1 Mice	0.25, 2.5, 25.0 and 250 µg/Kg bw/day	Perinatally and peripubertally	In addition to initial perinatal BPA exposure, the peripubertal BPA treatment aggravated adverse effects in female mice however, seemed to decrease changes in body weight and composition among control and BPA-treated males.	(Rubin <i>et al.</i> , 2017)
BPA, BPE and BPS	C/D-1 Mice	50 µg/Kg or 10 mg/Kg	From birth to PND 60	Treatment of mice with BPA and its analogues (BPE and BPS) disrupted progression of germ cell development. In female mice, postnatal BPA and BPE treatment enhanced onset of puberty and augmented body weight after parturition. Moreover, postnatal exposure of all the studied bisphenols increased steroid hormone concentrations in serum. Overall, BPA analogues affected male and female reproductive system.	(Shi <i>et al.</i> , 2017)
BPS	Zebrafish larvae	1, 3, 10 and 30 µg/L	From 2 to 168 hpf	BPS exposure altered whole-body THs and TSH levels as well as expression of key genes associated with HPT axis, therefore triggering thyroid-endocrine disruption.	(Zhang <i>et al.</i> , 2017)
BPS	Zebrafish ( <i>Danio rerio</i> )	1, 10 or 100 µg/L	28 days	BPS at environmentally relevant levels tend to impair glucose homeostasis in male zebrafish probably through hindering physiological effect of insulin furthermore, BPS at higher doses noticeably interfered with glucose-metabolism.	(Zhao <i>et al.</i> , 2018)

Dpf = Day-post-fertilization; GD = Gestational day; Hpf = Hour-post-fertilization); HPT = Hypothalamic-pituitary-thyroid; PND = Postnatal day; TH = Thyroid hormones; TSH = Thyroid stimulating hormone.

linings and plastic leachates (Porras *et al.*, 2014). Among bisphenols, median concentrations in urine specimens of individuals in US were detected as 0.72 (BPA), followed by 0.13 (BPS) and 0.08 ng/mL (BPF) (Zhou *et al.*, 2014). Human studies entailing bisphenols (BPF and BPS along with BPA) in different studied groups and their consequences on human health are summarized in Table 2. Asimakopoulos *et al.* (2015) analyzed urine specimens in general population from Saudi Arabia and found the concentrations of bisphenol analogues as 13.3 (BPS), 4.92 (BPA), 0.30 (BPAP), 0.19 (BPF), 0.05 ng/mL (BPB and BPAF).

A study comprising data from US and 7 Asian countries reported BPS in 81% of human urine samples (21 ng/mL) (Liao *et al.*, 2012b). Highest BPS levels were detected in urine samples collected from Japan as 1.180 followed by 0.299 (US) and 0.226 ng/mL (China and other countries). A number of bisphenol analogues were reported in urine specimens from Chinese people living nearby a BPAF production plant as 0.886 (BPA), 0.018 (BPAF), 0.228 (BPF) and 0.029 ng/mL (BPS) (Yang *et al.*, 2014 a, b).

In India, BPS concentration in children resulted as 0.05 (obese) and 0.61 (non-obese) ng/mL (Xue *et al.*, 2015). Moreover, BPS turn out to be most prominent analogue in urine specimens from people of Saudi Arabia with median concentration of 4.92 as compared to 2.01 (BPA) and 2.16 ng/mL (BPF) (Asimakopoulos *et al.*, 2015). Another study by Cobellis *et al.* (2009) demonstrated the occurrence of BPB in human sera (5.15 ng/mL) from Italian female that was larger as compared to 2.91 ng/mL (BPA).

As BPA is under regulations, human exposures to BPF and BPS is likely to increase therefore, bio-monitoring of bisphenol analogues is necessary. In this background, Zhou *et al.* (2014) detected bisphenol analogues and their median concentrations were observed as 0.72 (BPA), 0.13 (BPS) and 0.08 ng/mL (BPF). Peng *et al.* (2016) detected environmental estrogens including bisphenol A, nonylphenol and phthalate monoesters in urine samples of Chinese women (Nanjing area) with unexplained recurrent miscarriage. The average concentration of BPA in women with recurrent miscarriage was 7.13 ng/mL as compared to control group (4.43 ng/mL). On contrary, Yamamoto *et al.* (2016) found no relationship between BPA levels of maternal and fetal blood. During pregnancy, frequent

**Table 2.** Toxicological effects of BPA and bisphenol analogues in human studied groups

Target group	N	Biological samples	Country	Techniques	Analytes	LOD	Mean (range)	Effects	References
Pregnant women from R study	100	Urine	Netherlands	GC-MS/MS	BPA	0.26 µg/L	GM was 1.1 µg/L (<LOD-46.0 µg/L)	BPA was detected in 82% of analyzed samples.	(Ye <i>et al.</i> , 2008)
Pregnant women from MoBa	110	Urine	Norway	GC-MS/MS	BPA	0.26 µg/L	BPA GM conc. was 2.81 µg/L	Whether the higher BPA levels in pregnant women from MoBa were attributed to canned food (fish or seafood) consumption, though cannot be established from data. In addition, exposure from other than canned food and different environmental media like water and sediments might be taken into account.	(Ye <i>et al.</i> , 2009)
Children (aged 7-8 years)	127	Urine	Korea	HPLC-MS/MS	BPA	0.005 µg/L	GM conc. was found 1.02 µg/L	Elevated urinary concentrations of BPA in children (aged 7-8 years) were found to be positively related to wheezing and asthma	(Kim <i>et al.</i> , 2014)
Mother-child pairs	98	Urine	Sweden	LC-MS/MS	BPA	0.05 µg/L	GM conc. in mothers and children were as 1.31 and 1.48 0.05 µg/L, respectively	BPA was detectable (>LOD) in all urine samples. Urinary BPA levels were significantly associated between mothers and their children.	(Larsson <i>et al.</i> , 2014)
Pregnant women enrolled in P4 study	80	Urine	Canada	GC-MS/MS	BPA	0.2 µg/L	GM ranged 1.2-1.5 µg/L	Low reproducibility and sensitivity of BPA was found throughout pregnancy and post-partum period although, much greater reproducibility within a day. It was established that, for the average exposure within pregnancy, the maternal/fetal exposure assessment might be more correct if several measurements collected throughout the pregnancy as compared to only one spot measure.	(Fisher <i>et al.</i> , 2015)
Reproductive aged couples	501	Urine	USA	HPLC-MS/MS	BPA	0.02 ng/mL	Maternal and paternal urinary BPA conc. were 0.38 and 0.59 ng/mL, respectively	Preconception maternal as well as paternal urinary BPA concentrations might be associated with small birth size and greater gestational age however, the results seemed to be parent and chemical specific.	(Smarr <i>et al.</i> , 2015)
Men and women participants	3246	Urine	China	LC-MS/MS	BPA	0.30 ng/mL	Median urinary BPA conc. was 0.81 ng/mL	BPA exposure was negatively associated with hypertension and early macrovascular disorders among middle-aged and older Chinese population.	(Wang <i>et al.</i> , 2015)

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Target group	N	Biological samples	Country	Techniques	Analytes	LOD	Mean (range)	Effects	References
Adult women (A case-control study)	212	Urine	Cyprus and Romania	Tandem mass spectrometry	BPA	10 ng/L	Median BPA was found in Romanian (3778 ng/L) and Cypriot (1508 ng/L) population	Urinary BPA was observed to be positively associated ( $p < 0.05$ ) with serum TSH.	(Andrianou <i>et al.</i> , 2016)
Adult women (A case-control study)	212	Urine	Cyprus and Romania	Tandem mass spectrometry	BPF	13 ng/L	Median BPF in Romanian (416 ng/L) and Cypriot (485 ng/L) population	There was no relationship found between urinary BPF and TSH.	(Andrianou <i>et al.</i> , 2016)
Pregnant woman	30	Urine	Australia	LC-MS/MS	BPA	0.1 ng/mL	GM was found to be 5.0 ng/mL	Detection frequency of BPA was 100% in analyzed samples.	(Heffernan <i>et al.</i> , 2016)
Adult men and women	50	Urine	Brazil	LC-MS/MS	BPA	0.1 ng/mL	GM urinary BPA was found as 1.9 ng/mL	Overall, BPA was detected in 92% of the analyzed samples.	(Rocha <i>et al.</i> , 2016)
Pregnant women	245	Urine (first and second trimester)	USA	ID-HPLC-MS/MS	BPA	0.4 µg/L	SG-adj. GM was 1.39 µg/L (1 <sup>st</sup> trimester) and 1.27 µg/L (2 <sup>nd</sup> trimester)	There was no relation between BPA concentrations at first-trimester and glucose levels. But, second trimester BPA levels were positively associated with blood glucose levels implying that BPA exposure during second trimester can adversely affect glucose levels among sub-fertile female.	(Chiu <i>et al.</i> , 2017)
Primiparous women included in POPUP study	178	Urine	Sweden	LC-MS/MS	BPA	0.22 ng/mL	BPA mean 1.67 ng/mL (<LOD-15.9 ng/mL)	Significantly declining temporal trend was reported for urine concentration of BPA.	(Gyllenhammar <i>et al.</i> , 2017)
Primiparous women included in POPUP study	178	Urine	Sweden	LC-MS/MS	BPF	0.03 ng/mL	BPF mean was 0.12 ng/mL (<LOD-1.38 ng/mL)	BPF showed an ascending temporal trend implying that BPF is of no use to substitute BPA from health-risk viewpoint.	(Gyllenhammar <i>et al.</i> , 2017)
Primiparous women included in POPUP study	178	Urine	Sweden	LC-MS/MS	4,4-BPF	0.03 ng/mL	1.09 ng/mL (<LOD-24.3 ng/mL)	Interestingly, an ascending temporal trend was noticed for 4,4-BPF during the same study period.	(Gyllenhammar <i>et al.</i> , 2017)
Maternal/fetal pairs	30	Maternal urine and fetal cord blood (MUAFCB)	USA	UPLC/MS	BPS	0.5 ng/mL	Median urinary BPS conc. was 0.19 ng/mL	60% of tested maternal urine samples were positive for BPS. However, BPS was not detectable in cord blood.	(Ihde <i>et al.</i> , 2017)
Maternal/fetal pairs	30	MUAFCB	USA	UPLC/MS	BPA	1.7 ng/mL	Not indicated	7% of mothers tested positive for BPA while, it was found in one cord blood sample.	(Ihde <i>et al.</i> , 2017)

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Target group	N	Biological samples	Country	Techniques	Analytes	LOD	Mean (range)	Effects	References
Maternal/fetal pairs	30	MUAFCB	USA	UPLC/MS	BPB	0.5 ng/mL	Not indicated	57% of mothers tested positive for BPB in their urine. BPB was detected in only one cord blood sample.	(Ihde <i>et al.</i> , 2017)
Maternal/fetal pairs	30	MUAFCB	USA	UPLC/MS	BPE	1.9 ng/mL	Not indicated	BPE was not found in any maternal urine or fetal cord blood samples.	(Ihde <i>et al.</i> , 2017)
Maternal/fetal pairs	30	MUAFCB	USA	UPLC/MS	BPF	2.5 ng/mL	Not indicated	BPF was detectable in 50% cord blood and 17% maternal samples. In contrast to high pervasiveness of BPF in cord blood, the occurrence of this compound was low in maternal urine samples.	(Ihde <i>et al.</i> , 2017)
Maternal/fetal pairs	30	MUAFCB	USA	UPLC/MS	BPAF	0.14 ng/mL	Not indicated	BPAF was detected in 83% of tested maternal urine and 57% cord blood samples. The study proposed the prevalence of these compounds in environment as well as higher exposure among pregnant female and newborns.	(Ihde <i>et al.</i> , 2017)
Children with atopic dermatitis	18	Urine	Korea	HPLC-MS/MS	BPA	0.15 µg/L	3.36 µg/g Cr. In morning and 3.09 µg/g Cr. in afternoon	An association was observed between BPA exposure and aggravation of atopic dermatitis symptoms in children.	(Kim <i>et al.</i> , 2017)
Preschool children	113	Urine	Sweden	LC-MS/MS	BPA	0.09 µg/L	GM of urinary BPA was 1.4 µg/L	The estimated daily BPA exposure of children through ingestion of dust were lower than the reference values in most of preschools.	(Larsson <i>et al.</i> , 2017)
Preschool children	113	Urine	Sweden	LC-MS/MS	BPS	0.007 µg/L	GM of urinary BPS was 0.19 µg/L	BPS ranged 0.02-33.0 µg/L in urine samples of preschool children.	(Larsson <i>et al.</i> , 2017)
Preschool children	113	Urine	Sweden	LC-MS/MS	4,4 BPF	0.02 µg/L	GM of urinary 4,4 BPF was 0.16 µg/L	4,4 BPF ranged <LOD-32.0 µg/L in urine samples of preschool children.	(Larsson <i>et al.</i> , 2017)
Preschool children	113	Urine	Sweden	LC-MS/MS	2,2 BPF	0.01 µg/L	GM of urinary 2,2 BPF was 0.01 µg/L	2,2 BPF ranged <LOD-0.07 µg/L in urine samples of preschool children.	(Larsson <i>et al.</i> , 2017)
Adults (20 years and older) from NHANES	1521	Urine	USA	HPLC-MS/MS	BPA	0.2 µg/L	Median conc. in obese (1.5 ng/mL) and non-obese (1.1 ng/mL)	After adjusting demographic, socioeconomic, lifestyle parameters and urinary creatinine levels, BPA was found to be significantly related to general and abdominal obesity.	(Liu <i>et al.</i> , 2017)
Adults (20 years and older) from NHANES	1521	Urine	USA	HPLC-MS/MS	BPF	0.2 µg/L	Median conc. in obese (0.4 ng/mL) and non-obese (0.3 ng/mL)	On contrary, BPF was not associated with obesity in American adults at present exposure concentrations.	(Liu <i>et al.</i> , 2017)

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Target group	N	Biological samples	Country	Techniques	Analytes	LOD	Mean (range)	Effects	References
Adults (20 years and older) from NHANES	1521	Urine	USA	HPLC-MS/MS	BPS	0.1 µg/L	Median conc. in obese (0.4 ng/mL) and non-obese (0.3 ng/mL)	BPS was not related to obesity in American adults at present exposure concentrations.	(Liu <i>et al.</i> , 2017)
Mother-child pairs	470	Urine (first and third trimester)	Spain	LC-MS	BPA	0.1 µg/L	Maternal urinary GM (2.3 µg/L) and ranged 0.3-61.8 µg/L	No association was observed between BPA exposure and fetal growth.	(Casas <i>et al.</i> , 2016)
Participants recruited in ASNS	594	Urine	Austria	HPLC-MS/MS	BPA	1.25 µg/L	Total BPA ranged ND-17 µg/L	Generally, low BPA concentrations were observed among Austrian population. But, children and adolescents revealed higher exposure.	(Hartmann <i>et al.</i> , 2016)
Pregnant women	112	Urine	USA	LC-MS/MS	BPA	0.05 ng/mL	GM for total BPA was 7.69 ng/mL	Higher concentrations of BPA analytes might attributed to low-income of majority of individuals and direct analytic testing, which allows complete assessment of BPA exposure. A near-universal BPA exposure was observed among pregnant women raising other concerns for impacts on fetal development.	(Gerona <i>et al.</i> , 2016)
Women of reproductive age	305	Urine	Korea	HPLC-MS/MS	BPA	0.1 µg/L	BPA ranged from 0.1 to 18.3 µg/g Cr with mean value of 1.7 µg/g Cr	For Korean women, beverage consumption indicated a positive association with urinary BPA. Odds ratio for higher BPA concentration in female who had >100 g of beverage intake was significantly greater compared to those who consumed ≤100 g.	(Jo <i>et al.</i> , 2016)
Mothers and their children (follow-up) from MIREC study	812	Urine	Canada	GC-MS/MS	BPA	0.2 ng/mL	Overall BPA median (0.8) and mothers urinary BPA during pregnancy (<LOD-79.1 ng/mL)	86% of mothers had detectable BPA levels. Prenatal urinary BPA levels was found to be associated with some parameters of child (3 years age) behavior and some relations were stronger among boys.	(Braun <i>et al.</i> , 2017)
Mother-infant pairs in APPrON	132	Maternal urine (2 <sup>nd</sup> trimester)	Canada	HPLC-MS	BPA	0.32 ng/mL	GM in child sex: male (1.10 ng/mL) and female (1.04 ng/mL) with ranges 0.16-43.42 and 0.23-39.84 ng/mL, respectively	Similar to creatinine, the cortisol was detected in all samples but, two samples with values (> 3 µg/dL), which were not plausible biologically were removed. Prenatal BPA exposure is related to sex-specific changes in hypothalamic-pituitary-adrenal (HPA) axis function of infant.	(Giesbrecht <i>et al.</i> , 2017)

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Target group	N	Biological samples	Country	Techniques	Analytes	LOD	Mean (range)	Effects	References
Men	560	Urine	China	Modified HPLC	BPA	0.12 µg/L	GM was obtained as 0.44 µg/L and 0.50 µg/g Cr (adj.)	The detectable level of BPA was found in 70.4% of specimens. Hence, environmental BPA exposure was related to increased serum concentrations of LH and FSH among smokers, along with reduced serum levels of total testosterone in men with BMI $\geq 25$ kg/m <sup>2</sup> . The study suggested that impact of environmental exposure to BPA on hormonal levels could be altered by smoking and BMI. A positive relation was observed between urinary BPA levels and endometrioma.	(Liang <i>et al.</i> , 2017)
Women with endometrioma and control group	100	Urine	Iran	HPLC	BPA	0.33 ng/mL	GM of BPA in endometrioma group (3.14 ng/mL) and control group (0.72 ng/mL)		(Rashidi <i>et al.</i> , 2017)

2,2-BPF = 2,2-Bisphenol F; 4,4-BPF = 4,4-Bisphenol F; AD = Atopic dermatitis; APrON = Alberta Pregnancy Outcomes and Nutrition; ASNS = Austrian Study on Nutritional Status; Cr = Creatinine; Cr-adj. = Creatinine adjusted; GM = Geometric mean; HPA = Hypothalamic-pituitary-adrenal; LC-MS = Liquid chromatography-mass spectrometry; MIREC = Maternal Infant Research on Environmental Chemicals Study; MoBa = Norwegian Mother and Child Cohort Study; MUAFCB = Maternal urine and fetal cord blood; NHANES = The National Health and Nutrition Examination Survey; P4 study = Plastics and Personal-care Products use in Pregnancy (P4) study; POPUP study = Persistent Organic Pollutants in Uppsala Primiparas; R study = The Generation R Study; SG-adj. = Specific gravity adjusted; TSH = Thyroid stimulating hormone; UPLC/MS = Ultra performance liquid chromatography/mass spectrometry.

consumption of beef and pork was found to be positively correlated to maternal BPA concentrations. BPA level of maternal blood was determined as 0.419 whereas, in cord blood was found as 0.217 ng/mL.

In United States, Liu *et al.* (2017) observed relationship between obesity and bisphenols exposure in 1521 participants and reported that obese adults had greater levels of BPA, BPS and BPF as compared to non-obese individuals. After adjusting demographic, socioeconomic parameters and urinary creatinine levels, neither BPF nor BPS, but only BPA was found to be significantly related to obesity (general and abdominal) in US adults at present exposure concentration. Also, odds ratios (ORs) for general obesity were found as BPA (1.78), BPF (1.02) and BPS (1.22). Pell *et al.* (2017) examined the association between the parental concern regarding the environmental chemical contacts and urinary phthalate and phenol levels in their school-age children (8 years) in Ohio. They reported that concentrations of 4 phthalates, BPA and BPS were lower in children whose parents (n = 122) were more concerned about environmental chemicals than those (n = 96) who did not.

Owing to an increase volume of scientific studies and public knowledge regarding the adverse endpoints of BPA exposure, there is beginning to phase out BPA from numerous products. Previous studies have documented that being an endocrine disruptor, application of BPA has been inhibited in various products. As, several products are marketed with a “bisphenol A-free” guarantee due to its replacement with BPS however, these products are not “BPA-free”, and their usage might be associated with considerable reproductive toxicity (Glausiusz, 2014).

**Pakistan’s perspective.** On contrary, situation is even more worst in developing countries like Pakistan. BPA has not been recognized as a real concern at national level and there is limited data available that is however, based on scientific research. Due to insufficient scientific data regarding levels of BPA and its analogues in food matrices make the estimation of their dietary intake more problematic in Pakistan. In current scenario, production and marketing of bisphenols containing materials will tend to increase and so does health risks due to absence of regulation and governing bodies. In this milieu, a more systematic framework is required in Pakistan for regulating this ubiquitous chemical. It

is the need of hour that bio-monitoring studies of bisphenols should be conducted at national level to evaluate the health-related risk of Pakistani population particularly, infant, children and pregnant women. Pakistan Environment Protection Agency and Provincial Food Authorities should develop policies and regulatory framework and take stern on bisphenol compounds in food packaging and other food contact materials.

On the other side, environmental experts and health professionals should organize community programmes for public awareness for health menace of bisphenols. Such approaches could also be effective to determine association between bisphenol exposure and occurrence of hormone-based disorders. Nonetheless, human exposure to bisphenols is inevitable, it could be reduced by adopting some steps. Food and beverages contained in cans are the major contributor for bisphenol A toxicity thus, consumption of canned food should be reduced. Polycarbonate plastic packaging can also release bisphenol A on heating and at recurrent use therefore, the practice to microwave food in polycarbonate material should be restricted.

## Conclusion

Due to restriction on applications of bisphenol A in baby bottles and other packaging container, there has been an impetus to find safer BPA alternatives. Therefore, industry has switched from the removal of BPA towards the application of its substitutes. Published data reported negative endpoints of BPA analogues comparable or even higher than BPA. These contaminants are harmful to various organ and systems of human body including reproduction, thyroid and nervous system. Preferably, alternatives used in order to substitute a chemical must be inert, or less harmful as compared to original compound. Unfortunately, similar to BPA, its structural analogues have been under scrutiny due to their endocrine disrupting activities. Importantly, the use of bisphenol family of chemicals should be kept at minimum particularly for food contact materials. Moreover, the safety assessment of a certain chemical or chemical replacement should be carried out before it takes its way to market particularly, as substitutes for toxic compounds. There should be an adequate public awareness to encourage people particularly, children and adolescents to use packaging other than plastics such as glass, stainless steel, paper and ceramic.

**Conflict of Interest.** The authors declare no conflict of interest.

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